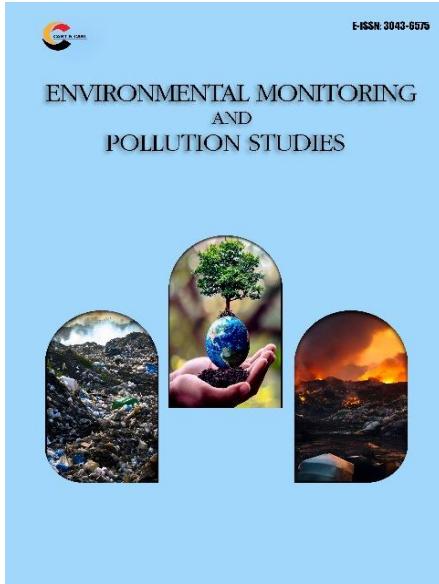




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Distribution and Levels of Contaminants of Emerging Concern in Surface Water of the Imo River, Nigeria: A Non- Targeted Analysis

Abstract

This is an assessment of the distribution and levels of contaminants of emerging concerns (CECs) in the surface water of Imo River, Nigeria, using a non-target approach. Three locations; Ekenobizi (Imo State), Owerri (Abia State), and Oyigbo (Rivers State) along the river were sampled based on high human density and significant anthropogenic activities. Surface water samples were collected from upstream, midstream, and downstream of each site using the grab sampling method and analysed in the laboratory within three hours of collection. A non-targeted screening was performed using gas chromatography-mass spectrometry (GC-MS) after sample preparation. The results indicated contamination of the river with eighty-five (85) unique CECs, distributed as follows: 21 CECs at Ekenobizi, 27 at Owerri, and 40 at Oyigbo. Some CECs were specific to particular locations, while others were detected across all three sampling sites. Among the identified compounds were siloxanes, fatty acids, amines, alkanes/alkenes, fluorinated organic compounds, nitrogenous compounds, and other organic chemicals. These chemicals can be traced back to pharmaceutical and personal care products, agro-chemicals, and various industrial chemicals, with each exhibiting different levels of ecological risk.

Keywords: Contaminants of Emerging Concern (CECs), Non-Targeted Analysis, Surface water, Imo River

Introduction

Many recent environmental studies have shown increasing presence of contaminants of emerging concern (CECs) in the environment. The aquatic environment is impacted significantly, raising global concerns due to potential adverse effects CECs could have on aquatic life, human health and the entire environment even at low concentrations (Sauve & Desrosiers, 2014; Yuange et al., 2024; Jianglu et al., 2024).

In a review of how one defines emerging contaminants preferably termed “contaminants of emerging concern” and what can be included in that group of contaminants, Sauve and Desrosiers, (2014) gave a historical perspective on the evolution of the issues surrounding emerging contaminants and how environmental scientists have tackled this issue. They noted that it began with global lead contamination from the Romans two millennia ago, moved on to arsenic-based and Dichlorodiphenyltrichloroethane (DDT) issues and more recently to pharmaceuticals, cyanotoxins, personal care products, nanoparticles, flame retardants, etc. Contaminants of emerging concern will remain a moving target as new chemical compounds are continuously produced and science continuously improves its understanding of current and past contaminants.

Contaminants of emerging concern (CECs) also referred to as emerging contaminants (ECs), or emerging pollutants (EPs), can be defined as newly identified synthetic or naturally occurring chemicals or biological agents that are detected in the environment and potentially hazardous or recently determined to be hazardous to humans and ecosystems. The risks associated with these contaminants are not fully understood, their toxicological significance is difficult to assess and generally accepted concentration limits for drinking water and discharge limits for wastewater effluent have not yet been established (Pal et al., 2014).

They may include pharmaceuticals and personal care products (PPCPs), per- and poly-fluoroalkyl substances (PFAS), endocrine disruptors (EDs), emerging pathogens, cyanotoxins and other natural toxins, pesticides, industrial chemicals, micro/nano plastics, nanomaterials, antibiotic resistance genes (ARGs), and other exogenous substances that are found in the environment but are not yet well understood in terms of their impacts on humans and natural ecosystems (Fang et al., 2024; Enyoh et al., 2020; Ebele et al., 2017). The occurrence of these contaminants in the aquatic environment and especially in surface water is a serious public health concern because potable drinking water and water for other uses are primarily sourced from surface and ground waters in many countries (Abafe et al., 2023; Galindo-Miranda et al., 2019)

Conventional water treatment processes were designed to tackle known inorganic and organic pollutants but with the increasing list of pollutants, treatment of water to potable standards have become increasingly difficult (Abafe et al., 2023). The analysis of these pollutants in water for treatment and management purposes is usually achieved through targeted analysis approach, using established and validated liquid chromatography, coupled with gas chromatography mass spectrometry detection techniques (Pitarch et al., 2010). However, these established techniques only cover a fraction of known pollutants, leaving many legacy and emerging compounds undetected and unstudied. Target analysis alone does not present the actual pollution status of the water systems (Abafe et al., 2023).

Non-targeted analysis (NTA) methods have unique value in that they can garner informative chemical measurements from samples of interest without the need for predefined chemical targets (McCord et al., 2022). The non-targeted analysis of environmental samples generally starts with the collection of samples to be analyzed, followed by analysis in which full scan mass spectrum (MS) data as well as tandem mass spectrum (MS/MS) fragmentation information is collected for identification purposes (Ng, 2021). After the data is acquired, one of the most important steps is the data pre-processing as it is required to make sense of the data and reduce not only the quantity, but also the complexity. This is done through a series of processes such as the detection of peaks, alignment of retention times,

background subtraction using blanks. The final step in the identification of compounds of interest by NTA involves utilizing all the information obtained from the previous steps, in which the MS and MSn data are matched up with their respective molecular ion, isotopic pattern and fragments (Hollender et al., 2017; Ng, 2021). In a Study conducted by Wu et al., (2023), a total of 70 CECs was detected at least once at selected seventeen (17) sampling sites along the Yangtze River. Twenty-four (24) CECs were detected at each site, and these were mostly pharmaceutical, personal care products and pesticides.

The Imo River located in south eastern Nigeria is a very important water resource that supports different human activities including, but not limited to fishing, agriculture, and domestic usage (Ogbonna et al., 2021). For the assessment of the occurrence and distribution of priority contaminants of emerging concern in Imo River, a non-targeted approach was employed to give insights into the actual CEC pollution level of the river.

Materials and Methods

Study Area

The Imo River, located in Southeastern Nigeria originates from the Imo State, and flows through Abia, Rivers and Akwa ibom states of Nigeria into the Atlantic ocean (Ogbonna et al., 2021). According to the location/topographic map of the study region, the Imo River Basin is located between latitudes 4° 38'N and 6° 01'N and between longitudes 6° 53'E and 7° 32'E. It has an approximate area of 9100 km² (Uma, 1989). Three (3) sampling locations along the Imo river were selected for the study based on high human density with much anthropogenic activities. The three locations were, Ekenobizi (Imo State), Owerrinta (Abia State) and Oyigbo (Rivers State) (Figure 1). The coordinates for each of the sampled points were as follows; Ekenobizi upstream (Lat 5.558823°, Long. 7.413593°), midstream (Lat 5.558852°, Long. 7.414024°), downstream (Lat 5.558812°, Long. 7.414057°), Owerrinta upstream (Lat 5.30586, Long. 7.28721), midstream (Lat 5.30599, Long. 7.28721), downstream (Lat 5.30776, Long. 7.28748) and Oyigbo upstream (Lat 4.89035, Long. 7.14274), midstream (Lat 4.88812, Long. 7.14351), downstream (Lat 4.88566, Long. 7.14792).

The activities around the chosen locations as observed during reconnaissance surveys include dredging and collection of sharp sand, waste disposal and fishing common to the three locations. Other activities observed were farming activities, a major market that sold various household items and building materials, including various grinding mills at Ekenobizi area, animal

sacrifices at the Owerrinta area and illegal petroleum exploration and a major abattoir at the Oyigbo area.

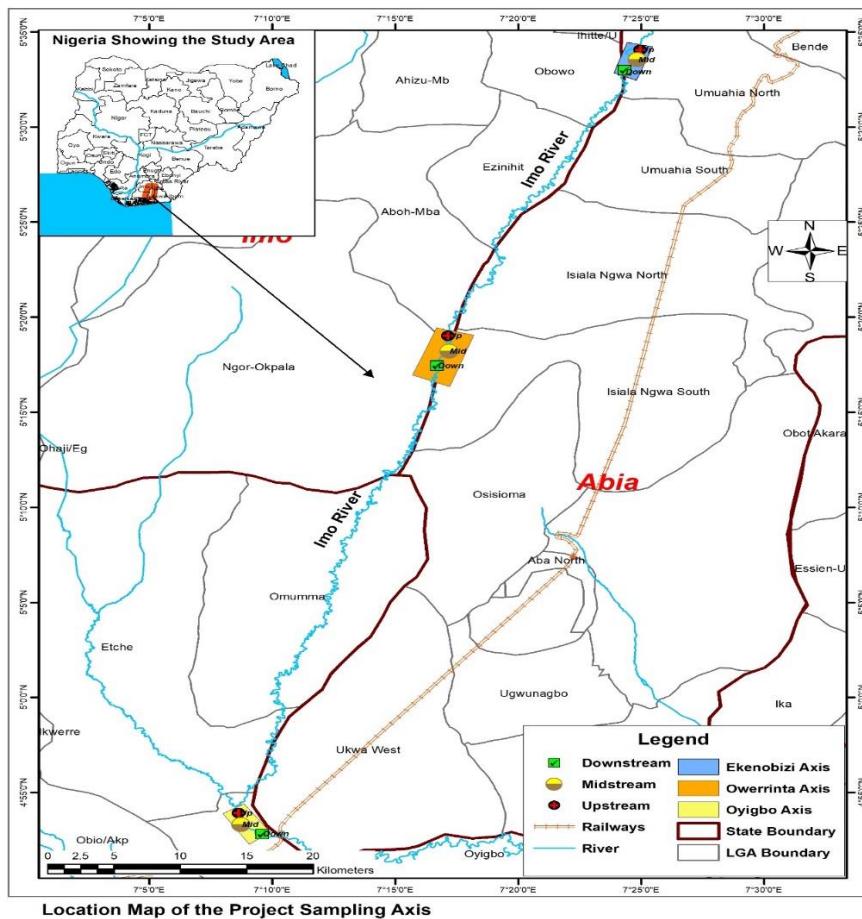


Figure 1: Overview of the Study Area and Sampled Locations

Sampling

Surface water samples were collected from the upstream, midstream and downstream parts of the river making three samples from each of the sampling locations (Ekenobizi, Owerrinta and Oyigbo respectively) using the grab sampling method (APHA, 2017). The help of local fishermen was employed at each of the locations to collect the samples using new and clean plastic bottles at a depth of 0.1 – 0.5 meters. The plastic bottles were thoroughly rinsed with the river water at every sampling point to avoid interference from unwanted impurities. The samples were collected directly into the bottles and filled to the brim leaving no space and covered tightly. Samples were moved to the lab for testing immediately after collection from each of the locations within three hours. A total of nine (9) samples were collected, three from each of the sampled locations.

Laboratory Analysis

The method used for the laboratory analysis was the Gas Chromatography/Mass Spectrometry (GC/MS) scan as described in Sunkara et al (2025). The GC/MS combines the separation powers of gas chromatography with the detection capabilities of mass spectrometry to enhance the efficiency of sample studies. This technique is well-recognised for its ability in unknown compound analysis. It enables the analysis of a broad range of compounds including organic acids, amino acids, sugars, fatty acids, sterols, and various xenobiotics (Sunkara et al, 2025).

The surface water samples were prepared as follows: 500ml of each of the surface water samples was introduced into a separating funnel, alongside equal volume of organic solvent (methanol) as the extracting solvent. The mixture was agitated and vent intermittently to release pressure. The agitation lasted for 15 minutes before allowing the two immiscible liquids to separate into layers. The stopper of the separating funnel was removed and the organic layer was released into the beaker through a glass funnel

stuffed with glass wool and covered with sodium anhydrous sulphate. The extract was allowed to concentrate in the fume hood and packed in the GC vial at exactly 1ml concentration.

The GC works on the principle that a mixture will separate into individual substances when heated. The heated gases are carried through a column with an inert gas (such as helium). As the separated substances emerge from the column opening, they flow into the MS. Mass spectrometry identifies compounds by the mass of the analyte molecule. A library of known mass spectra, covering several thousands of compounds, is stored on a computer. Mass spectrometry detector identifies these analytes and match them with library. Identification is based on the molecular structure, molecular mass and calculated fragments elucidated by the MS Quadrupole mass analyser filtered on a mass to charge basis by the HED (High energy diode). They are qualitatively interpreted with aid of National Institute standard and Technology (NIST) spectrum database NIST 08 model. The name, molecular weight and structure of the components of the material are ascertained by use of the library. The relative percentage amount of each component was obtained by comparing its average peak area to the total area, the spectrum of the unknown components are compared to the 2008 version, through which the various spectra extractions and interpretation are obtained.

The Conc (mg/l) = % amount \times 0.002 \times 1000 (recall that 500ml of surface water samples were used for extraction)

Statistical analysis

The study adopted tabular representation of findings and the use of a chart to summarize and describe findings for

clarity. The hypothesis of the study was formulated and tested using the Analysis of Variance (ANOVA) at 95% level of significant. As inferential statistics, ANOVA allows for exploring the statistically significant differences between two or more variances. One-ANOVA was adopted for the hypotheses testing and it is expressed as thus;

$$F = \frac{MST}{MSE}$$

$$MST = \frac{\sum_{i=1}^k (T_i^2/n_i) - G^2/n}{k-1}$$

$$MSE = \frac{\sum_{i=1}^k \sum_{j=1}^{n_i} Y_{ij}^2 - \sum_{i=1}^k (T_{i=1}^2/n_i)}{n-k}$$

Where;

F= Variance ratio of the overall test

MST= Mean Square due to treatment /groups (between groups)

MSE= Mean square due to error (within groups, residual mean square)

Y_{ij}= an Observation

T_i= Group total

G= Grand total of all observations

n_i= number in the group _i and n is total number of observations

Results and Discussion

The outcome of the non-targeted scan for CECs in the surface water across the sampled axes of the Imo River is presented in Figure 2 and Tables 1 to 3. A total of eighty-five (85) CECs were detected which detailed as 21 CECs at Ekenobizi axis (23%), 27 CECs at Owerriinta axis (32%) and 40 CECs at Oyigbo axis (45%) across the upper stream (US), midstream (MS) and downstream (DS) points of each location.

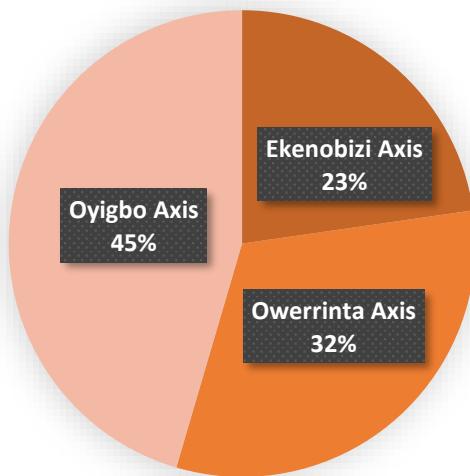


Figure 2: Total CECs in Surface Water across the Sampled Axes of the Imo River

At Ekenobizi axis (Table 1), Acetaldehyde (C_2H_4O) was detected at 7.64 mg/L at the upstream (U.S) point, but not detected downstream, Ethylene oxide (C_2H_4O) ranged from 7.214 mg/L (U.S) to 8.13 mg/L (MS), Eicosamethyl cyclodecasiloxane ($C_{20}H_{60}O_{10}Si_{10}$) recorded 7.76 mg/L (U.S), rising to 18.96 mg/L (M.S) and 19.02 mg/L (D.S), Ethyl 2-((diethoxyphosphoryl)oxy)-3,3,3 trifluoropropanoate peaked at 18.04 mg/L (D.S). Several compounds like 2-methyl Adenosine, 1-Heptadecanamine, Hexadecanoic acid derivatives were location-specific, ranging from 5.76 mg/L to 14.58 mg/L. Octadecamethyl cyclononasiloxane showed high concentration in both US (12.18 mg/L) and D.S (12.34 mg/L), Among all the CECs, Eicosamethyl

cyclodecasiloxane ($C_{20}H_{60}O_{10}Si_{10}$), Ethyl 2-((diethoxyphosphoryl)oxy)- 3,3,3 trifluoropropanoate ($C_9H_{16}F_3O_6P$) and Octadecamethyl cyclononasiloxane ($C_{18}H_{54}O_9Si_9$) were detected across US, MS and DS indicating possible persistence in flow. Similarly, some of the CECs were detected across two of the sampling locations. At Ekenobizi (Table 1) and Owerrinta (Table 2) axes, Hexasiloxane, tetradecamethyl- was detected at 12.36mg/l and 5.98mg/l respectively, 1,1,1,5,7,7,7-Heptamethyl-3,3- bis(trimethylsiloxy) tetrasiloxane was also detected at 10.2mg/l and 160.24mg/l respectively at both locations. Eicosane was also detected from 2.03mg/l concentration at Owerrinta, (Table 2) up to 20.52mg/l concentration at Oyigbo (Table 3).

Table 1: CEC Concentration in the Surface Water at Ekenobizi Axis of Imo River

Name Of Compound	Formulae	Surface Water (mg/l)		
		US	MS	DS
1 Acetaldehyde	C_2H_4O	7.64	ND	ND
2 Ethylene oxide	C_2H_4O	7.214	8.13	ND
3 Eicosamethyl cyclodecasiloxane, -	$C_{20}H_{60}O_{10}Si_{10}$	7.76	18.96	19.02
4 Ethyl 2-((diethoxyphosphoryl)oxy)- 3,3,3 trifluoropropanoate	$C_9H_{16}F_3O_6P$	11.9	15.74	18.04
5 2-methyl Adenosine, -	$C_{11}H_{15}N_5O_4$	ND	6.09	7.38
6 1-Heptadecanamine	$C_{17}H_{37}N$	ND	10.81	12.31
7 Hexadecanoic acid,2,2,3,3,4,4,5,5 ,6,6,7,7 dodecafluoroheptyl ester	$C_{23}H_{34}F_{12}O_2$	ND	12.32	14.58
8 4,6-dimethoxy-5-nitro- Pyrimidine	$C_6H_7N_3O_4$	10.18	ND	ND
9 Hex-5-enylamine	$C_6H_{13}N$	9.14	ND	ND
10 Octadecamethyl cyclononasiloxane, -	$C_{18}H_{54}O_9Si_9$	12.18	5.97	12.34
11 1-Dodecanamine	$C_{12}H_{27}N$	ND	5.76	ND
12 1,1,1,5,7,7,7-Heptamethyl-3,3- bis(trimethylsiloxy)tetrasiloxane	$C_{13}H_{40}O_5Si_6$	10.2	ND	ND
13 L-Alanine, 3-[(aminocarbonyl)amino]-	$C_4H_9N_3O_3$	7.54	4.66	ND
14 Silane,[[4-[1,2-bis[(trimethylsiloxy]ethyl]- 1,2phenylene]bis(oxy)] bis[trimethyl-	$C_{20}H_{42}O_4Si_4$	6.56	ND	ND
15 3,5,5-trimethyl 1-Hexanamine	$C_9H_{21}N$	9.84	ND	ND
16 Hexasiloxane, tetradecamethyl-	$C_{14}H_{42}O_5Si_6$	12.36	ND	ND
17 18-Nonadecen-1-amine	$C_{19}H_{39}N$	ND	13.48	15.10
18 Phthalic acid, butyl hexyl ester	$C_{18}H_{26}O_4$	ND	6.52	ND
19 3-Butyn-1-ol	C_4H_6O	ND	ND	9.50
20 N-[1-[(butylchloroboryl)oxy]-2 chloroethylidene]-2- chloro Ethanimidamide	$C_8H_{14}BCl_3N_2O$	ND	22.38	21.78
21 Decyl 3,5-dimethylphenyl ester Phthalic acid	$C_{26}H_{50}O_4$	ND	ND	7.52

US = Upstream, MS = Midstream, DS = Downstream, ND = Not detected

At Owerrinta Axis (Table 2), 1,1,1,5,7,7,7-Heptamethyl...tetrasiloxane ($C_{13}H_{40}O_5Si_6$) recorded an exceptionally high value of 160.24 mg/l (US), 6-Octadecenoic acid ($C_{18}H_{34}O_2$) showed 152.6 mg/L (MS), Decanedioic acid bis(2-ethylhexyl) ester ($C_{26}H_{50}O_4$)

peaked at 146.3 mg/l (DS). Lower concentrations included cis-Vaccenic acid (15.4 mg/l US), Oleic Acid (21.2 mg/l DS), Eicosane (2.25 mg/l DS), and others in sub-1.0 mg/L ranges like Triacontyl pentafluoropropionate (0.06 mg/l at US) and 17-

Pentatriacontene (0.01 mg/l at US). Decanedioic acid, bis(2-ethylhexyl) ester ($C_{26}H_{50}O_4$) was detected across US, MS and DS.

At Oyigbo Axis (Table 3), Pentadecane, 2,6,10,14-tetramethyl was most abundant at 19.82 mg/l (US), with 15.86 mg/l at DS. High concentration of CECs was detected with Hexadecane (19.16 mg/l US), Nonadecane

(19.62 mg/l US), and Eicosane (20.52 mg/kg US). Tridecane, Tetradecane, Undecane, and Methylene Chloride were also present in trace amounts (≤ 1.66 mg/l). A significant outlier was Trimethyl(4-tert-butylphenoxy)silane at 180.68 mg/l (M.S). Among all the CECs, Eicosane ($C_{20}H_{42}$), Heneicosane ($C_{21}H_{44}$) and Pentacosane ($C_{25}H_{52}$) were detected across US, MS and DS.

Table 2: CEC Concentration in the Surface Water Samples at Owerrinta Axis of Imo River

Name Of Compound	Formulae	Surface Water Conc. (mg/l)		
		US	MS	DS
1 n-Hexadecanoic acid	$C_{16}H_{32}O_2$	ND	8.82	7.13
2 1,1,1,5,7,7,7-Heptamethyl-3,3-bis(trimethylsiloxy)tetrasiloxane	$C_{13}H_{40}O_5Si_6$	160.24	ND	ND
3 6-Octadecenoic acid, (Z)-	$C_{18}H_{34}O_2$	ND	152.6	ND
4 Decanedioic acid, bis(2-ethylhexyl) ester	$C_{26}H_{50}O_4$	12.92	10.4	146.3
5 cis-Vaccenic acid	$C_{18}H_{34}O_2$	15.4	ND	4.42
6 Oleic Acid	$C_{18}H_{34}O_2$	ND	3.58	21.2
7 Eicosane	$C_{20}H_{42}$	ND	2.03	2.25
8 Terephthalic acid, di(2-ethylhexyl) ester	$C_{24}H_{38}O_4$	ND	1.38	2.18
9 Tetrapentacontane, 1,54-dibromo-	$C_{54}H_{108}Br_2$	0.14	1.92	ND
10 Hexasiloxane, tetradecamethyl-	$C_{14}H_{42}O_5Si_6$	5.98	ND	ND
11 Cyclohexane, 1-(1,5-dimethylhexyl)-4-(4-methylpentyl)-	$C_{20}H_{40}$	ND	2.54	ND
12 Octacosyl heptafluorobutyrate	$C_{32}H_{57}F_7O_2$	0.06	1.16	ND
13 trans-13-Octadecenoic acid	$C_{18}H_{34}O_2$	ND	1.44	ND
14 Ethanol, 2-(hexadecyloxy)-	$C_{18}H_{38}O_2$	ND	3.74	ND
15 p-Menth-8(10)-en-9-ol, cis-	$C_{10}H_{18}O$	ND	4.36	ND
16 Hexadeca-2,4,15-trienoic acid, ethyl ester	$C_{18}H_{30}O_2$	0.2	ND	ND
17 Triacontyl pentafluoropropionate	$C_{33}H_{61}F_5O_2$	0.06	ND	ND
18 Bromoacetic acid, octadecyl ester	$C_{20}H_{39}BrO_2$	0.02	ND	ND
19 17-Pentatriacontene	$C_{35}H_{70}$	0.00	ND	ND
20 2-Benzo[1,3]dioxol-5-yl-8-methoxy-3-nitro-2H-chromene	$C_{17}H_{13}NO_6$	0.28	ND	ND
21 Cyclononasiloxane, octadecamethyl-	$C_{18}H_{54}O_9Si_9$	2.10	ND	ND
22 2-Hexadecene, 3,7,11,15-tetramethyl-1-, [R-[R*,R*-(E)]]-	$C_{20}H_{40}$	0.8	ND	ND
23 Trisiloxane, 1,1,1,5,5-hexamethyl-3,3-bis[(trimethylsilyl)oxy]-	$C_{12}H_{36}O_4Si_5$	1.78	ND	ND
24 p-Menth-8(10)-en-9-ol, cis-	$C_{10}H_{18}O$	ND	ND	0.074
25 Oxirane, decyl-	$C_{12}H_{24}O$	ND	ND	0.09
26 Hexatriacontyl pentafluoropropionate	$C_{39}H_{73}F_5O_2$	ND	ND	0.62
27 9-Octadecenal, (Z)-	$C_{18}H_{34}O$	ND	ND	6.14

US = Upstream, MS = Midstream, DS = Downstream, ND = Not detected

Table 3: CEC Concentration in the Surface Water Samples at Oyigbo Axis of Imo River

Name of Compound	Formulae	Surface Water (mg/kg)		
		US	MS	DS
1 Pentadecane, 2,6,10,14-tetramethyl	C ₁₉ H ₄₀	19.82	ND	15.86
2 Hexadecane, 2,6,10,14-tetramethyl-	C ₂₀ H ₄₂	19.16	ND	14.18
3 Dodecane, 2,6,11-trimethyl-	C ₁₅ H ₃₂	0.14	ND	1.5
4 2-Bromo dodecane	C ₁₂ H ₂₅ Br	7.44	ND	5.88
5 Nonadecane	C ₁₉ H ₄₀	19.62	ND	13.34
6 Eicosane	C ₂₀ H ₄₂	20.52	0.58	14.58
7 Heneicosane	C ₂₁ H ₄₄	20.42	0.74	14.9
8 Docosane	C ₂₂ H ₄₆	18.84	ND	16.66
9 Tricosane	C ₂₃ H ₄₈	15.52	3.16	13.09
10 Tetracosane	C ₂₄ H ₅₀	10.74	ND	18.7
11 Dodecane	C ₁₂ H ₂₆	ND	ND	0.16
12 Undecane, 2,6-dimethyl-	C ₁₃ H ₂₈	ND	ND	0.2
13 Tridecane, 7-methyl-	C ₁₄ H ₃₀	ND	ND	0.44
14 Tridecane	C ₁₃ H ₂₈	ND	ND	1.66
15 Tetradecane	C ₁₄ H ₃₀	ND	ND	1.54
16 Decahydro-4,4,8,9,10 pentamethylnaphthalene	C ₁₅ H ₂₈	0.42	ND	1.46
17 Undecane	C ₁₁ H ₂₄	ND	ND	3.42
18 Pentadecane	C ₁₅ H ₃₂	2.68	ND	8.64
19 Hexadecane	C ₁₆ H ₃₄	4.72	ND	4.96
20 Pentacosane	C ₂₅ H ₅₂	10.66	0.48	17.60
21 Hexacosane	C ₂₆ H ₅₄	ND	ND	13.52
22 Heptacosane, 1-chloro-	C ₂₇ H ₅₅ Cl	2.46	ND	2.44
23 Eicosane, 10-heptyl-10-octyl-	C ₃₅ H ₇₂	ND	ND	1.26
24 17.alpha.(H),21.beta.(H)-Hopane	C ₃₅ H ₇₂	0.94	ND	0.86
25 Docosyl pentafluoropropionate	C ₂₅ H ₄₅ F ₅ O ₂	0.08	ND	ND
26 2,6,10-Trimethyltridecane	C ₁₆ H ₃₄	2.14	ND	ND
27 Tetradecane, 2,6,10-trimethyl-	C ₁₇ H ₃₆	2.84	ND	ND
28 Heptadecane, 3-methyl-	C ₁₈ H ₃₈	2.62	ND	ND
29 Heptacosane	C ₂₇ H ₅₆	4.14	ND	ND
33 Octacosane	C ₂₈ H ₅₈	0.62	ND	ND
31 28-Nor-17.alpha.(H)-hopane	C ₂₉ H ₅₀	0.86	ND	ND
32 Methylene Chloride	CH ₂ Cl ₂	ND	0.57	ND
33 Octadecane, 3-ethyl-5-(2-ethylbutyl)-	C ₂₃ H ₄₈	ND	0.50	ND
34 Benzo[h]quinoline, 2,4-dimethyl-	C ₁₅ H ₁₃ N	ND	0.16	ND
35 Terephthalic acid, 2-ethylhexyl octyl ester	C ₂₄ H ₃₈ O ₄	ND	0.34	ND
36 Octadecane	C ₁₈ H ₃₈	ND	0.28	ND
37 Decanedioic acid, bis(2-ethylhexyl) ester	C ₂₆ H ₅₀ O ₄	ND	10.94	ND
38 Silicic acid, diethyl bis(trimethylsilyl) ester	C ₁₀ H ₂₈ O ₄ Si ₃	ND	0.8	ND
39 Corydaldine	C ₁₁ H ₁₃ NO ₃	ND	0.18	ND
40 Trimethyl(4-tert.-butylphenoxy)silane	C ₁₃ H ₂₂ OSi	ND	180.68	ND

US = Upstream, MS = Midstream, DS = Downstream, ND = Not detected

Table 4 shows the parent groups of the detected CECs and their percentages in relation to the total number of CECs detected in each location. At Ekenobizi, the water samples showed higher percentage of Siloxanes/Silanes and Amines/Amides at 23.8% for each of both groups followed by Fatty acids/esters/derivatives at 14.3% and Alkanes/Alkenes at 9.5%. The rest of the groups were at

4.8% each and the miscellaneous group ranked 9.5%. At Owerrinta, the water samples had Fatty acids/esters and derivatives at 33.3%, Alkanes/Alkenes at 18.5%, Siloxanes/silanes at 14.8%, Fluorinated organic compounds and Alcohols/glycols at 11.1% each, polycyclic and Aromatic group at 3.7% and the miscellaneous group at 7.4%. At Oyigbo, the water

samples showed a high content of Alkanes/Alkenes at 72.5%. Siloxanes/ Silanes, Fatty acids/ Esters/Derivatives and the Polycyclic/Aromatics groups

all contributed 5% each; Fluorinated organic compounds, 2.5% and the Miscellaneous group 10%.

Table 4: Percentage Parent Groups Classification of Detected Chemicals of Emerging Concern in the Imo River Surface Water

Group	Ekenobizi	Owerrinta	Oyigbo
Siloxanes and Silanes	23.8%	14.8%	5%
Fatty acids, Esters and Derivatives	14.3%	33.3%	5%
Alkanes and Alkenes	9.5%	18.5%	72.5%
Amines and Amides	23.8%	ND	ND
Fluorinated Organic Compounds	4.8%	11.1%	2.5%
Nitrogenous bases, Amino Acids and Peptides	4.8%	ND	ND
Polycyclic and Aromatic Compounds	4.8%	3.7%	5%
Alcohols and Glycols	4.8%	11.1%	ND
Pesticides and related organometallics	4.8%	ND	ND
Other groups / Miscellaneous	9.5%	7.4%	10%
TOTAL	100%	100%	100%

ND – Not detected

Research Hypothesis

From Table 5, the study's hypothesis was tested using the ANOVA. The hypothesis was tested based on the following statement:

H_0 : There is no significant difference in the CECs concentrations in the surface water of samples along the Imo River

H_1 : There is a significant difference in the CECs concentrations in the surface water of samples along the Imo River

Based on the outcome, the null hypothesis (H_0) which stated that there is no significant difference in the CECs concentrations in the surface water of samples along the Imo River was not rejected (where $p > 0.05$, $p = 0.992$).

Table 5: Significant Different Analysis of CECs in Surface Water

		Sum of Squares	Df	Mean Square	F	Sig.
Surface Water	Between Groups	0.577	2	0.289	0.008	0.992
	Within Groups	2164.355	57	37.971		
	Total	2164.932	59			

Previous studies have indicated that the aquatic environments. are receptacles for domestic and industrial wastes (Abafe et al., 2023), therefore, surface water screening supports monitoring, management and sustainability of the aquatic system (Afolabi et al., 2024). The use of non-target analysis (NTA) procedure for this study has been previously adopted by the study conducted by Tian et al. (2019), Penalver et al. (2022), and Abafe et al. (2023). The screening of Imo river using

gas chromatography (GCMS) was able to establish the presence of eighty-five (85) unique CECs in the surface water. This is similar to the findings of Wu et al., (2023) who detected a total of 70 CECs in the surface water at the selected 17 sampling sites of the Yangtze River.

The widespread detection of siloxanes in surface water, particularly at concentrations exceeding 160 mg/L, reflects the increasing use of personal care products,

lubricants, and industrial materials, consistent with global reports of siloxane pollution in aquatic environments (Mojstiewicz-Pieńkowska et al., 2016). Siloxanes, commonly used in personal care products and industrial applications, are known for their resistance to degradation, potential for bioaccumulation, and endocrine-disrupting effects (Wang et al., 2013). The elevated levels detected, particularly near populated or industrial zones, suggest anthropogenic discharges, consistent with reports of siloxane contamination near urban water bodies (Horii & Kannan, 2008).

Fatty acids, such as Hexadecanoic acid and cis-Vaccenic acid, detected at significant concentrations, have been previously associated with anthropogenic input from domestic, industrial and agricultural sources (Rudel et al., 2011, Meng et al., 2020). The detection of fluorinated fatty acid derivatives like *Hexadecanoic acid, dodecafluoroheptyl ester* raises concerns due to their association with perfluoroalkyl substances (PFAS). PFAS are globally recognized for their environmental persistence and bioaccumulation, posing risks to aquatic life and human health (Giesy & Kannan, 2002). Their detection in Imo River samples reflects patterns seen in polluted rivers worldwide, such as the Danube and Rhine (Ahrens et al., 2010). High detection of fluorinated organics, including *Triacontyl pentafluoropropionate* and *Octacosyl heptafluorobutyrate*, reflects global concerns over PFAS-related compounds. PFAS contamination has been linked to adverse health outcomes in fish, such as liver damage and hormonal disruptions (De Silva et al., 2021). The findings corroborate similar studies in West African aquatic systems, where PFAS residues have been reported at concerning levels, necessitating urgent monitoring (Olutona et al., 2023).

The widespread presence of linear and branched alkanes (*Pentadecane, Heptacosane, Eicosane*), alongside PAH derivatives and biomarkers like *Hopanes*, suggests inputs from petroleum-related activities, urban runoff, or atmospheric deposition. Studies have shown that aliphatic hydrocarbons and hopanes are reliable indicators of oil pollution and crude oil contamination in aquatic environments (Wang et al., 2004). This aligns with the results especially from the Oyigbo axis (72.5% alkanes/alkenes) where illegal petroleum exploration was observed. Additionally, PAHs, known for their mutagenic and carcinogenic properties, pose significant ecological threats (Afolabi et al., 2024). Primary long-chain amines (*Nonadecylamine, Dodecanamine*) and nitrogenous amides suggest sources from agrochemical formulations, detergents, or industrial effluents. Long-chain amines can exhibit aquatic toxicity and contribute

to bioaccumulation in fish (Poste et al., 2014). The presence of nitrogen-containing organic pollutants in aquatic systems is increasingly documented, with evidence linking such compounds to developmental toxicity in aquatic organisms.

The identification of halogenated compounds such as *Bromoacetic acid esters, Methylene Chloride*, and various brominated alkanes suggests both industrial discharge and degradation by-products from chlorination processes. Such compounds are persistent, bioaccumulative, and often exhibit toxicity to aquatic life (USEPA, 2018). Halogenated hydrocarbons are established indicators of anthropogenic pollution, especially in industrialized coastal regions (UNEP, 2019). Detection of nitrogenous heterocycles (*Pyrimidine derivatives, Indole compounds*) and aromatic amines points to pharmaceutical and agricultural runoff, consistent with recent research linking these classes to antimicrobial resistance hotspots in aquatic ecosystems (Kümmerer et al., 2018). The co-occurrence of persistent siloxanes, Fluorinated organics, PAHs, alkanes/alkenes and Nitrogenous pollutants implies a complex mixture of contaminants, elevating risks of synergistic toxicity, bioaccumulation, and trophic transfer. Similar mixtures have been linked to endocrine disruption, reproductive impairment, and immune suppression in aquatic organisms (Ankley et al., 2005).

The Anova result which showed no significant difference among the CEC concentrations across the sampled locations suggests that the factors influencing the CEC contamination of the Imo River are similar and consistent enough across the sampled locations that they produce no measurable difference in the CEC levels. This is consistent with the similar anthropogenic presence and activities at the various locations.

Conclusion

The non-target analysis (NTA) of surface water from the Imo River at Ekenobizi, Owerrinta and Oyigbo using GCMS was able to establish the presence of various contaminant compounds of emerging concerns. A total of 85 unique CECs was detected. The detected compounds were of diverse parent/functional groups with diverse chemical characteristics.

The detection and concentration patterns of Contaminants of Emerging Concern (CECs) in the Imo River surface water reveal critical trends of environmental contamination and potential ecological risks, consistent with global studies. The findings underscore significant contamination of the Imo River

system with a broad spectrum of emerging organic pollutants, including siloxanes, fatty acids, amines/amides, alkanes/alkenes and fluorinated compounds. Findings resonate with global research on the persistence, mobility, and ecological risks of these contaminants.

1. Targeted analysis of detected CECs for the purpose of setting permissible health limits and standards for treatment and management purposes.
2. Implementation of stricter monitoring and control of industrial effluents, especially from manufacturing, petroleum, and personal care product industries, which are known sources of siloxanes, amines, and fatty acids.
3. Promotion of sustainable farming practices to reduce the discharge of fatty acids and nutrient-rich compounds into the river, as supported by global mitigation strategies.
4. Establishment of a long-term, monitoring framework targeting CECs in surface water, following international guidelines for surface water quality preservation.
5. There is need for public health awareness in form of consumption advisories for surface waters and to discourage household waste disposal into the river for communities around the Imo River.
6. Replacement of high-risk compounds, particularly persistent siloxanes and fluorinated esters, with biodegradable alternatives, aligning with global green chemistry initiatives.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Credit Authorship Contribution Statement

Humphrey-Ebenezer, J. N: Conceptualization, Methodology, Formal analysis, Investigation, Resources, Data curation, Visualization, Project administration, Writing - original draft. **Nwakanma, C. & Nzegbule, E.:** Supervision, Methodology, Validation, Formal analysis, Data curation, Visualization, Review & Editing.

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